

Synergistic effects of climate and land-use change on representation of African bats in priority conservation areas

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Abstract

Bats are considered important bioindicators and deliver key ecosystem services to humans. However, it is not clear how the individual and combined effects of climate change and land-use change will affect their conservation in the future. We used a spatial conservation prioritization framework to determine future shifts in the priority areas for the conservation of 169 bat species under projected climate and land-use change scenarios across Africa. Specifically, we modelled species distribution models under four different climate change scenarios at the 2050 horizon. We used land-use change scenarios within the spatial conservation prioritization framework to assess habitat quality in areas where bats may shift their distributions. Overall, bats' representation within already existing protected areas in Africa was low (~5% of their suitable habitat in protected areas which cover ~7% of Africa). Accounting for future land-use change resulted in the largest shift in spatial priority areas for conservation actions, and species representation within priority areas for conservation actions decreased by ~9%. A large proportion of spatial conservation priorities will shift from forested areas with little disturbance under present conditions to agricultural areas in the future. Planning land use to reduce impacts on bats in priority areas outside protected areas where bats will be shifting their ranges in the future is crucial to enhance their conservation and maintain the important ecosystem services they provide to humans.

Keywords: Land-use Change, Climate Change, Chiroptera, Conservation, Zonation, Africa, Connectivity

1. Introduction

Current rates of species extinctions and declines are unprecedented (Butchart *et al.*, 2010). Species extinction rates are now 1,000 times higher than the ‘background’ rate (De Vos *et al.*, 2014). Across vertebrates, 16 to 33% of species are considered to be globally threatened (Hoffmann *et al.*, 2010). This biodiversity ‘crisis’ is driven by anthropogenic factors, such as overutilization of species, habitat destruction, pollution and the introduction of invasive species (Diamond, 1984). Human-induced climate change is also expected to affect species persistence into the future (Bellard *et al.*, 2012). While the original cause of decline may be driven by one anthropogenic factor, extinction is often driven by multiple interacting pressures (Brook *et al.*, 2008). Hence, more studies are urgently needed to unveil the combined effects of extinction drivers on biodiversity.

Spatial conservation prioritization deals with the identification of priority areas where limited resources should be allocated for conservation actions (Moilanen *et al.*, 2009). An important goal of spatial conservation prioritization deals with the identification of areas where threats will impact biodiversity (Margules and Pressey 2000). However, conservation planning assessments often ignore the dynamic nature of threats (Moilanen *et al.*, 2009). Particularly, conservation planners should anticipate the rates and patterns of dynamic threats, such as future climate change and land-use change (Krauss *et al.*, 2010; Bellard *et al.*, 2012). Quantitative scenarios can be used to evaluate the impact of future socio-economic development pathways on biodiversity and ecosystem services to optimize current conservation actions to reduce future biodiversity loss (Pereira *et al.*, 2010).

Bats (Order Chiroptera) constitute about 23% of mammal diversity (Wilson and Reeder, 2005). Further, bats provide important ecosystem services such as predation of insects (Kalka *et al.*, 2008), seed dispersal (Shilton *et al.*, 1999) and pollination (Sazima *et al.*, 1989). They are also important bioindicators (Jones *et al.*, 2009, Mehra *et al.*, 2011), as well as indicators of specific human impacts, such as the quality of water courses (Scott *et al.*, 2010). Globally bats are under threat from anthropogenic activities, such as agricultural and urban expansion and over-utilisation of resources (Voigt and Kingston, 2016). Arguably, the highest threat to bats can be ascribed to widespread changes in land use systems predicted in the future (Hannah *et al.* 1995, Verburg *et al.*, 2013), because these will impact bat fitness directly in terms of roost, food and water loss. At the same time, climate change will shift bats' ranges and likely drive species to local extinction (Rebelo *et al.*, 2010). However, it is less clear how synergistic effects of climate and land use change will affect priority areas for the conservation of bats (Hughes *et al.*, 2012).

Compared to other continents, African bats have been poorly monitored and knowledge of their distribution is still very scarce (Martin *et al.*, 2013). More than 30% of African bat species are classified as threatened or data deficient (IUCN, 2014). Our goal was to identify the spatial priorities for the conservation of 169 bat species under expected climate and land use change across Africa. Our objectives were (i) to model bats' distributions under present and future climatic conditions; (ii) assess bats' representation inside already existing protected areas; and (iii) identify priorities for conservation actions outside protected areas under future climate and land use change.

2. Materials and methods

2.1 Biodiversity features

Recorded locations for 169 bat species (14,050 total records with the number of locations per species ranging from 5 to 604) of the approximately 250 species in Africa were used to model species distributions (SDMs) for the African continent (Appendix Table A.1, Appendix Figure A.1). Because occurrence data were based on museum records and we lacked absence data we used MaxEnt to estimate the distribution for each bat species (Phillips *et al.*, 2006). Among the many options for building SDMs from presence-only species records (Latimer *et al.*, 2006; Thuiller *et al.*, 2009; Renner and Warton, 2013), MaxEnt has good predictive performance (Elith *et al.*, 2006; Radosavljevic and Anderson, 2014).

We downloaded 19 current climate variables from the World Clim for present climate and future climate in 2050 (Hijmans *et al.*, 2005). We performed a principal components analysis on these variables to control for autocorrelation between variables (Garcia *et al.*, 2012; Schoeman *et al.* 2013). Variables with the largest eigenvalues associated with the principal component axes were extracted ($n = 10$ variables), and compared using a correlation matrix. For pairs with $r > 0.8$, the variable with the higher eigenvalue was kept. Six variables were used in the SDMs: Temperature Seasonality, Minimum Temperature of Coldest Month, Mean Temperature of Warmest Quarter, Precipitation of Driest Quarter, Precipitation of Warmest Quarter, and Precipitation of Coldest Quarter. Future climate variables were derived from the Intergovernmental Panel on Climate Change (IPCC) AR5 scenarios RCP2.6 (Van Vuuren *et al.*, 2007), RCP4.5 (Smith and Wigley, 2006; Clarke *et al.*, 2007; Wise *et al.*, 2009), RPC6.0 (Fujino *et al.*, 2006; Hijioka *et al.*, 2008) and RPC8.5 (Riahi *et al.*, 2007).

SDM performance was evaluated with tenfold cross-validation, by partitioning data into two subsamples: one for calibration-validation and the other for evaluation (Hastie *et al.*, 2005). The average distribution likelihood from 10 repetitions was used as an input for the spatial conservation prioritization (see below). The predictive value of the likelihood maps were calculated against the test data using the receiver operator curve (ROC) value, which produces an area under the curve (AUC) value (Appendix Table A.2). The AUC value indicates the discriminatory value of the models to predict the likelihood of a presence point being higher than the likelihood of a pseudo-absence point (Phillips *et al.*, 2006). SDMs were run at 5 arc-minute resolution, which was also the resolution of the spatial conservation prioritization (see below) and suggested default settings for MaxEnt (Phillips and Dudík, 2008). We produced raster grids of the standard deviation for each species to be used in the prioritization analyses (Phillips and Dudík, 2006).

2.2 Spatial conservation prioritization

We implemented spatial priority ranking with the Zonation (v4) methods and software (Moilanen *et al.*, 2005; Lehtomäki and Moilanen, 2013; Di Minin *et al.*, 2014; Moilanen *et al.*, 2014) to identify the priority areas for the conservation of bats under future land use and climate change scenarios. As output, Zonation produces priority rank maps and corresponding performance curves, which describe how well represented each feature entered into the analysis is in any given top or bottom fraction of the priority map (landscape). The ranking balances all factors - including species distribution, connectivity, and possible costs - entered into the analysis. The additive-benefit function cell removal rule for aggregation of conservation value was used (see Moilanen *et al.*, 2011). The additive-benefit function computes a maximum-utility type solution, where value is additive across biodiversity features, and where feature-specific representation is converted to value via concave power

functions, which most commonly are parameterized according to the canonical species-area curve (Moilanen, 2007).

Appendix Figure A.2 shows a flowchart of analysis and data inputs used in Zonation. In Zonation, weights assigned to features influence the balance among features in the prioritization solution. Typically, weights have positive values, but can also be set to 0.0 in surrogacy analyses (Di Minin and Moilanen 2014), or even have negative values, for example when multiple opportunity costs are included in the analysis (Moilanen *et al.* 2011). In this study, species were weighted according to their current International Union for Conservation of Nature and Natural Resources (IUCN) Red List assessment: Least Concern (weight = 1), Near Threatened (weight = 2), Vulnerable (weight = 3) and Endangered (weight = 4) (Appendix Table A.1). As a precautionary measure, species that were Data Deficient were included with the same weight as vulnerable species in the analyses (Butchart and Bird, 2010).

To account for connectivity and the scale of landscape use of bat species, we induced aggregation by using distribution smoothing on species distribution grids (Moilanen *et al.*, 2014). Distribution smoothing is a species-specific aggregation method that emphasizes areas that are well connected to others, thereby resulting in a prioritization with more compact priority areas (Moilanen *et al.*, 2014). The smoothing effectively identifies important semi-continuous regions where the species has high levels of occurrence. In contrast, scattered occurrences in fragmented habitat lose relative priority. The connectivity of cells is determined with a smoothing kernel, where the radius of the kernel was approximated as the radius of the mean dispersal distances. In the analysis, we used the mean dispersal distances

for three functional groups of bats to calculate the parameter of a dispersal kernel for each species. Moilanen *et al.* (2014) provide full details on how to calculate the parameter of the dispersal kernel based on dispersal distances. The average dispersal distance for each functional group of bats was estimated as: (i) open-air bats with long and narrow wings adapted for energetically efficient flight over long distances (Molossidae, Rhinopomatidae, Emballonuridae) - 40 km (Best *et al.*, 2003); clutter-edge bats with wings of intermediate length and area adapted for flight at the edges of vegetation (Vespertilionidae, Miniopteridae) - 25 km (Jacobs *et al.*, 2005); and clutter bats with short and broad wings adapted for slow manoeuvrable flight in cluttered space (Rhinolophidae, Hipposideridae, Nycteridae, Pteropodidae, Megadermatidae) - 5 km (Fenton, 1983; Fenton *et al.*, 1985).

A number of studies have found the thematic resolution of the dynamic land use change scenarios to be insufficient for an appropriate representation of habitat categories for many ecological applications, especially at fine spatial resolutions (Barbet-Massin *et al.*, 2011, Martin *et al.*, 2013). Particularly, land use categories that closely relate to the species-specific habitat requirements are amalgamated in broader land use systems that are less relevant to the species distribution. In addition, land use change scenarios may not adequately capture gradual changes in species distribution, and may contain classification errors (Cord *et al.* 2014). As such, we decided to use land-use change scenarios within the spatial conservation prioritization framework and not as predictor variables in species distribution models. In Zonation, we used global land-use change models developed independently from this study (van Asselen and Verburg, 2013) as condition layers. The condition layer represents information about the quality and degradation of land use categories and its influence on biodiversity features or groups of features (see e.g. Di Minin *et al.*, 2016; Montesino Pouzols *et al.*, 2014).

In the land-use change models, land-use changes are driven by macro-economic assessment of regional demand and supply of agricultural commodities (Kram and Stehfest, 2012), which accounts for local factors that either promote or constrain land-use change (Verburg *et al.*, 2013). The land-cover change scenario was based on the OECD Environmental Outlook to 2050 (OECD, 2012). Land availability, as well as socio-economic and biophysical conditions, steer the model to convert land use systems, either resulting into an expansion of human dominated land use systems over (semi-)natural systems, or leading to an intensification of land management to fulfill world-region scale demands. At the same time, the model simulates abandonment of agricultural practices and re-wilding. As a result, the model provides a good representation of the multiple drivers of habitat loss.

To better interpret results, we simplified the 29 original land use types (van Asselen and Verburg, 2013) into six distinct land use types that we considered to have similar habitat suitability for bats: (i) intensive and extensive agriculture; (ii) agriculture with natural mosaic; (iii) forest and low intensity agriculture; (iv) grassland and low intensity agriculture; (v) bare ground; and (vi) urban areas (Appendix Table A.3, Appendix Figure A.3). Present (2000) and future (2050) land use maps were converted to numerical data by assigning the six land use types condition values between 0 and 1. Condition values were assigned according to disturbance of natural habitat and intensity of use (Appendix Table A.4). We created condition layers for each functional bat group (Appendix Table A.1). To reduce uncertainty on the impact that land use change will have on the three functional bat groups, we accounted for both an optimistic and a pessimistic scenario. Under both scenarios, condition values were highest for forest and low intensity agriculture, land use (value of 1.0) and lowest for bare

ground (Appendix Table A.4). The optimistic scenario assumed that human land use would have less of a negative impact on bat species. Under the pessimistic scenario, instead, agricultural and natural mosaics, grassland and low intensity agriculture, intensive and extensive agriculture, and urban were assumed to be more detrimental to bat species (with decreases of 0.3, 0.2, 0.3 and 0.2 in condition values respectively compared to the optimistic scenario) (Appendix Table A.4). The values in the original SDMs were then multiplied by the condition values for present and future (optimistic and pessimistic) scenarios, respectively, in Zonation. In the Zonation analyses, we used the transformed sets of SDMs for the present and future (optimistic and pessimistic) scenarios, respectively.

2.2.1 Shifts in spatial conservation priorities

In the analyses where climate change was considered, we used the connectivity interaction feature in Zonation, which allows the calculation of connectivity between the present and future distributions of a species (Carroll *et al.*, 2010; Kujala *et al.*, 2013; Moilanen *et al.*, 2014). Specifically, two connectivity distributions were used for each species. The connectivity from the current to the future distributions was used to consider dispersal from source areas to future distribution areas. The connectivity from the future to the present distributions was used to account for stepping-stones, which are expected to help species reach the core areas of their future distributions. The spatial scale of connectivity was set to the mean dispersal distances for the three functional groups of bats, as explained above. Thus, for each species we used four potentially relevant distribution models in spatial priority setting (present distribution, future distribution, dispersal source and stepping stone).

2.2.2 Accounting for uncertainty

We accounted for various sources of uncertainty in the analyses. First, we used the standard deviation grids for each SDM in an approach called distribution discounting to give the highest value to those species occurrences that have high mean likelihood and low standard deviation (Moilanen *et al.*, 2014). Distribution discounting was applied separately to the present and future modelled species distribution models before calculating connectivity between present and future layers. In addition, we accounted for the impacts of different climate scenarios (Appendix Figure A.4). Overall, a total of 18 scenarios (two for present optimistic and pessimistic land-use change scenarios; four optimistic and four pessimistic land-use change scenarios (for present conditions) for each IPCC scenario; and four optimistic and four pessimistic land-use change scenarios (for 2050 conditions) for each IPCC scenario) were developed (details are provided in Appendix Figure A.2). Finally, we averaged optimistic and pessimistic scenarios (for both present and future – 2050 – land use) into three main consensus scenarios (present; future climate with present land use; and future climate with future land use) (see e.g. Struebig *et al.*, 2015) (Appendix Figure A.2).

Under each scenario, we carried out a gap analysis in order to understand what the representation levels for bats were inside already existing protected areas and then identified priority areas for bats' conservation action outside already existing protected areas. We did so by using a hierarchical mask that identified the locations of extant protected areas at a scale of 5 arc-minutes. This way it is guaranteed that the highest priorities are located in existing protected areas (Lehtomäki and Moilanen 2013). We extracted the protected areas from the World Database on Protected Areas (<http://www.protectedplanet.net>) by selecting only protected areas belonging to IUCN protected area categories I to VI to focus on areas with 'formal' protection. To define the core areas for bat conservation outside already existing

protected areas we used the 17 percent of terrestrial land under protection promoted by Aichi target 11 of the Convention on Biological Diversity (Convention Biological Diversity, 2010; Tittensor *et al.*, 2014; Di Minin and Toivonen, 2015). We also compared the changes in area overlap of the top 17% priority selected cells between the different scenarios and between the different scenarios and land use types in ArcMap (ESRI, 2011).

3. Results

Priority areas for the conservation of bats outside existing protected areas fell in high species richness areas close to the equator (Fig. 1, Appendix Figure B.1). There was a greater shift between present and future priority areas when future land-use change was included in the analysis (Fig. 1, Appendix Table B.1). At the 17% priority areas, the overlap between present (Fig. 1A) and future climate change with present land use (Fig. 1B) was 81%, whereas the overlap between present (Fig. 1A) with future climate change and future land use (Fig. 1C) decreased to 63% (Appendix Table B.1). Under present land use and future climate (Fig. 1B), western and eastern African countries lose conservation priority areas to central African countries, such as the Central African Republic and The Democratic Republic of Congo. When both future land use change and climate change were included in the analysis (Fig. 1C), some priority areas shifted to the southern African region, specifically Angola, Zimbabwe and Zambia.

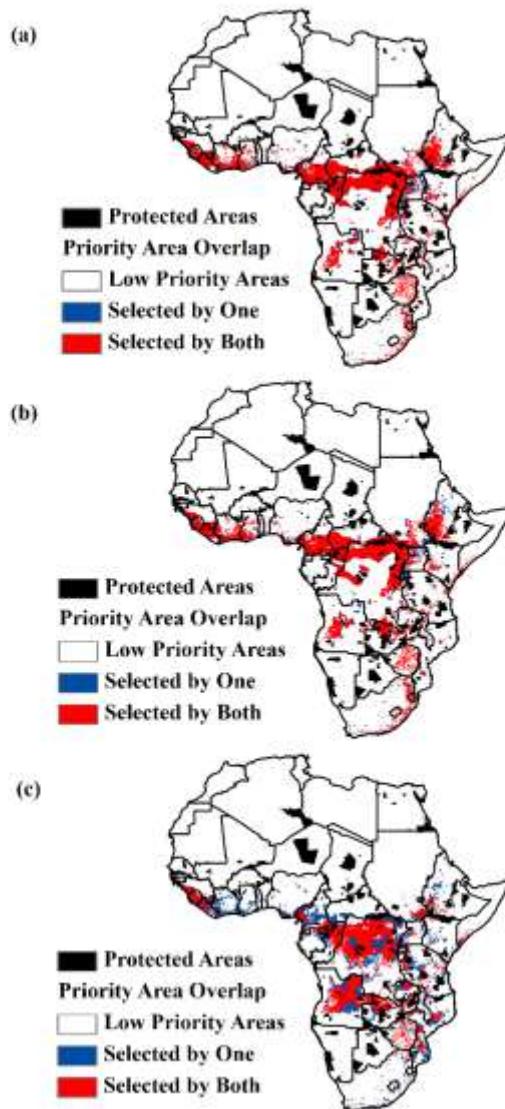


Figure 1. Overlap of spatial conservation priorities for bat conservation in Africa under (a) current climate and current optimistic and pessimistic land use scenarios; (b) future climate change and current optimistic and pessimistic land use scenarios; and (c) future climate and future optimistic and pessimistic land use scenarios. Areas within the 17% priority which were selected by both optimistic and pessimistic land use scenarios are in red and areas covered by either the optimistic or the pessimistic scenario are shown in blue. The priority rank maps for future climate change and current land use scenarios (b) and future climate and future land use scenarios (c) were averaged in a consensus scenario from four climate change scenarios for both the optimistic and the pessimistic land use scenarios.

The performance curves, which have direct correspondence to the priority rank maps, show that bat species are currently poorly represented (~ 5% on average) in existing protected areas (Fig. 2a, Appendix Table B.2). Importantly, the representation will decrease to 3% in the future because of the predicted shift in species distributions under climate change (Fig. 2b). Within the 17% top priority areas for conservation actions, the mean representation of bats was the lowest under future land use and climate change (Figs 2c, 3). At the 17% top priority areas for conservation action, there was a drop in representation of ~9% between present optimistic land use scenarios (Fig. 2a) and future climate and future pessimistic land use scenarios (Fig. 2c) for all functional bat groups.

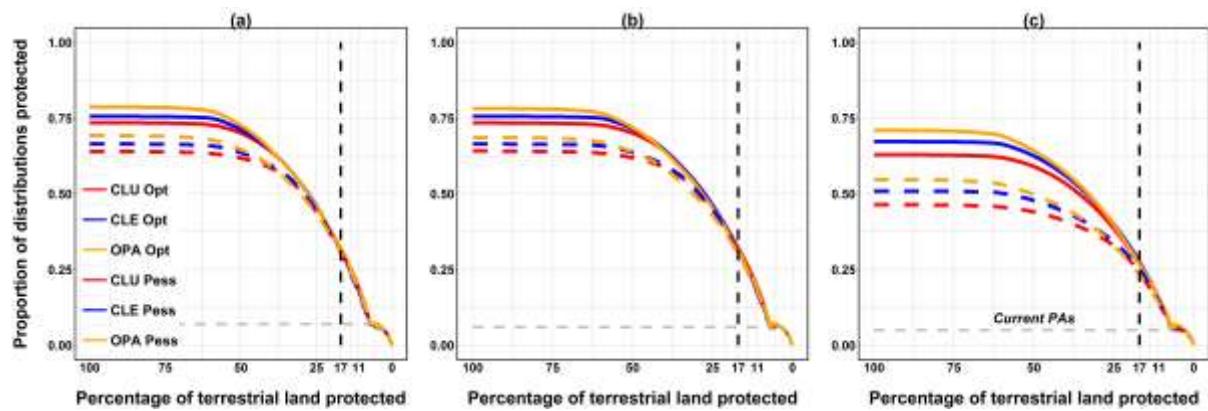


Figure 2. Performance curves, which quantify the mean proportion of the original distributions of three groups of bats in Africa at the 17% target for conservation, under (a) current climate and current optimistic and pessimistic land use scenarios; (b) future climate change and current optimistic and pessimistic land use scenarios; and (c) future climate and future optimistic and pessimistic land use scenarios. Open-air (OPA) = bats with long and narrow wings adapted for energetically efficient flight over long distances; clutter-edge (CLE) = bats with wings of intermediate length and area adapted for flight at the edges of vegetation; and clutter (CLU) = bats with short and broad wings adapted for slow manoeuvrable flight in cluttered space. Optimistic (opt) are the performance curves for optimistic land use scenarios and pessimistic (pess) are the performance curves for pessimistic land use scenarios. The horizontal dashed line in grey shows representation within current protected areas. The vertical line in black shows the 17% land target for conservation.

At the 17% top priority areas for conservation actions, the median, as well as the minimum and maximum, representations for clutter species was the lowest among the three functional groups of bats considered in the analyses (Figure 3). Particularly, the representation for clutter species was the lowest under the future land use and climate change scenarios (Figure 3). The representation for endangered species was the lowest under the future climate and future land use change scenarios (Appendix Table B.3).

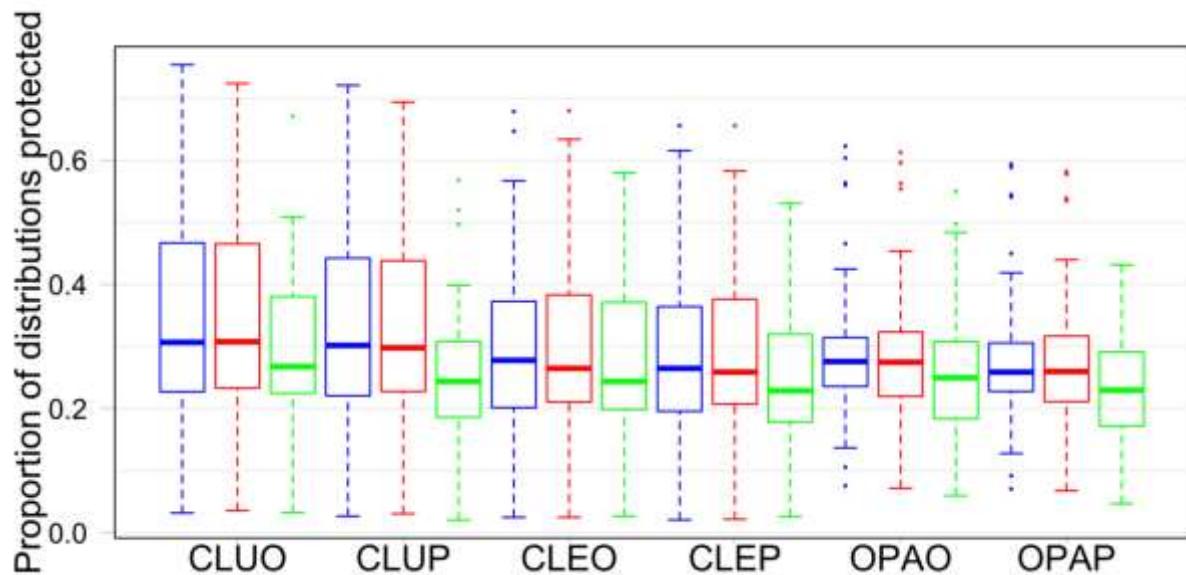


Figure 3. Box-and-whisker plot of representation levels for all bat species within the 17% target for conservation. The blue boxes represent current climate and current optimistic and pessimistic land use scenarios; red boxes represent future climate change and current optimistic and pessimistic land use scenarios; and the blue boxes represent future climate and future optimistic and pessimistic land use scenarios. CLUO stands for clutter under optimistic land use scenario, CLUP stands for clutter under pessimistic land use scenario, CLEO are clutter edge species under optimistic land use scenarios, CLEP are clutter edge species under pessimistic land use scenarios, OPAO are open air species with optimistic land use scenario and OPAP are open air species under pessimistic land use scenarios.

Finally, we found a shift in the land use classes represented within the 17% priority areas from the present to the future conditions (Table 1). Under present conditions, a large proportion of priorities were found in forested areas with low intensity agriculture. Under future land use, agricultural and natural mosaics and grasslands with low intensity agriculture will become more prevalent within the priority areas (Table 1).

Table 1. Percentage of the 17 percent priority areas within each land use type. The percentage was calculated for scenarios CCCO (current climate and current optimistic land use), CCCP (current climate and current pessimistic land use), FCCO (future climate and current optimistic land use), FCPO (future climate and current pessimistic land use), FCFO (future climate and future optimistic land use), FCFP (future climate future and future pessimistic land use)

	CCCO	CCCP	FCCO	FCCP	FCFO	FCFP
Intensive and extensive agriculture	1	1	8	7	3	1
Agriculture and natural mosaic	8	3	28	29	25	12
Forest and light use	69	74	26	28	42	55
Grassland and light use	8	8	22	21	14	13
Bare ground	14	14	14	15	15	18
Urban	0	0	1	1	0	0

4. Discussion

In this study, we identified priority areas for the conservation of bats in Africa. Our assessment for Africa is timely, as bats are a poorly monitored group and under increased threat because of overexploitation and persecution (Monadjem *et al.*, 2007). The conservation priority areas identified around the equatorial regions of Africa in this study correspond with the six central African hotspots and five of the six endemism hotspots of bat diversity in Africa (Herkt *et al.* 2016). Our results show an additional priority conservation area in central Angola near the hills and slopes on the east of the elevated Planato region to north-west Zambia and the southern tip of the Democratic Republic of Congo. Thus, similar to many

other biodiversity conservation prioritization studies (see e.g. Myers *et al.*, 2000), biodiversity-rich areas are concentrated towards the tropics (Gaston, 2000; Willig *et al.*, 2003). We found that incorporating future land-use change in combination with climate change had a greater effect on changing the location of future conservation priority areas than those with present land use only (Hampe and Petit, 2005; Schmitz *et al.*, 2015). In addition, our results illustrate that future land-use change will reduce species' representation both in existing protected areas and in areas identified as priorities for conservation actions. Finally, we predict a decrease in the proportion of priority areas falling in natural forest and an increase in the proportion of priority areas falling in agricultural mosaics. This likely poses the greatest challenge for bat conservation in Africa.

Only few studies included land-use change scenarios when predicting shifts in priority areas for conservation action (see e.g. Struebig *et al.*, 2015). Many studies suggest that climate change is a stronger driver of change in species distribution than land-use change (e.g. Martin *et al.*, 2013). However, Hughes *et al.* (2012) found a more marked retraction in projected distributions in Southeast Asian bats when including future vegetation changes. Indeed, our results confirm the importance of accounting for land use change when identifying priority areas for conservation action. Our results also highlight the importance of preventing further forest loss in central Africa to enhance bat conservation (Dirizo *et al.*, 2003). To better elucidate the synergistic effects of climate change and land use change, further studies should examine responses of bat species and assemblages to anthropogenic changes at finer scales (Kalda *et al.*, 2015). The challenge is to link broad-scale assessments, as we did here, to finer scale land use practices (Margules and Pressey, 2000), and to integrate this information into local area conservation plans and assessments (Keith *et al.*, 2008).

Taking dispersal ability at the functional level into account, we found that clutter and clutter edge species are most influenced by future human land uses (Fig. 2), probably because both functional groups require woody vegetation structures to survive. Similarly, vulnerable species in this analysis comprised mainly clutter (3 species, *Rhinolophus ruwenzorii*, *Rhinolophus guineensis* and *Hipposideros marisae*) and clutter-edge (2 species, *Myotis bocagii* and *Mops trevori*) species, with only one open air species (*Taphozous hildegardeae*), and their representation in priority areas decreased under the future land use scenario. Although high mobility of bats enables them to utilise natural habitat patches even in human-dominated urban and agricultural landscapes (Duchamp and Swihart, 2008, Davidai *et al.* 2015), bat species exhibit high variability in dispersal abilities at the mesoscale, even within functional groups (Meyer *et al.* 2008; Taylor *et al.* 2012). Thus, our models could be markedly refined by using data on dispersal ability for individual species. Regardless, shifts in current versus future priority areas highlight the need for movement links and stepping stones between conservation priority areas at broad spatial scales across country borders (Williams *et al.*, 2005; Schmitz *et al.*, 2015). Active monitoring of bat species in all land use types could play an important role in identifying where corridors should be enhanced or preserved (Medellín *et al.*, 2000).

Using target groups of indicator species to determine the status of biodiversity is an integral tool to assess and plan conservation strategies. These species should be sensitive to changes and sampled efficiently as well as yield objective results (Moreno *et al.*, 2007). Bats play a number of key biological functions within ecosystems and represent an important group for biodiversity themselves (Wilson and Reeder, 2005). They also have life history traits such as low reproductive output and long life expectancy which render them effective bioindicators

of the state of ecological systems, pertinently with regard to changes in climate, land use and habitat quality (Jones *et al.*, 2009; Cunto and Bernard, 2012, Heer *et al.*, 2015). Presumably, bats are similarly useful bioindicators in biodiversity planning as typical surrogates such as mammals and birds (Di Minin and Moilanen, 2014; Rodriguez and Brooks, 2007; Jones *et al.* 2009). However, few studies have tested this, for example the correlation of bat responses to habitat change with those of other surrogate species (Rodriguez and Brooks, 2007). Thus, although bats may have great potential as bioindicators, other indicator taxa that exploit the landscape in similar ways should also be included in monitoring surveys (Pocock and Jennings, 2008).

In conclusion, our analyses identify priority areas for bat conservation by considering global change and accounting for connectivity using estimates of functional group dispersal. The combined effects of future land-use and climate changes will have marked effects on the distributions and thus conservation priority areas of African bats. We highlight the importance to preserve high value forest patches in central Africa. Our results show that areas where bats are likely to move following climate change might already be transformed to land uses that do not support bats. Although verified data on bat occurrence, particularly from Africa, are often lacking from open source databases (IUCN, 2015), predicting future distributional shifts in response to climate change and land use allocation, as we did here, is key to identifying priority areas for conservation actions under dynamic threats (Myers *et al.*, 2000). What is now needed are innovative conservation efforts at fine spatial scales within land use categories. These include efforts to stop habitat clearing, mitigate threats such as culling and utilisation of bats as bushmeat, and encourage maintenance or rehabilitation of natural vegetation to enhance connectivity (Voigt and Kingston, 2016). To increase the awareness of bats to urban planners we suggest routinely including bats in ecosystem service

assessments (Kunz *et al.*, 2011). Some bat species may do well inside of transformed landscapes for example urban exploiters and adapters (Jung and Kalko 2011, Schoeman 2015, Jung and Threllfall 2016). In the African context, conservation strategies in priority areas that include human development are key (Lele *et al.*, 2010). These strategies could, for example, involve community based education initiatives that focus on ecosystem services provided by bats (e.g. Lele *et al.*, 2010; Boyles *et al.* 2011).

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Appendix A

Supplementary Methods

Table A.1 List of species used in the analyses Records from southern Africa: Schoeman et al. 2013. Records from East Africa: Swynnerton and Hayman (1951); Aellen (1957); Hayman and Hill (1966); Verschuren (1967); Niort (1970); Anciaux de Faveaux (1972); Verschuren (1980); Kock (1981); Aggundey and Schlitter (1984); Schlitter and Aggundey (1986); Baeten et al. (1984); Kock and Howell (1988); Montfort (1992); Heller et al. (1994); Fahr et al. (2002); Stanley and Kock (2004); Stanley et al. (2005a); Stanley et al. (2005b); Stanley et al. (2005c); Stanley et al. (1996); Thorn et al. (2007); Bergmans (1997); Rovero et al. (2007); Stanley and Foley (2008); Stanley (2009); Kock and Stanley (2009); Kock et al. (2000); Thorn and Kerbis Peterhans (2009); and a visit to the Mammal Department, National Museums of Kenya, Nairobi (by AM). Records from the Horn of Africa: Funaioli and Lanza (1968); Kock (1969); Largen et al. (1974); Hill (1975); Koopman (1975); McLellan (1986); Hill et al. (1988); Kock and Howell (1988); Varty and Hill (1988); Duckworth et al. (1993); Yalden et al. (1996); Kruskop and Lavrenchenko (2000); Kruskop and Lavrenchenko (2006); Pearch et al. (2001); Fenton and Eger (2002); Benda et al. (2004); Lavrenchenko et al. (2004); and Bergmans (1988). Records from Central Africa: Hayman et al. (1966); Aellen (1952); Lunde and Beresford (1997); van Cakenberghe et al. (1999); Cotterill (2001); Eger and Schlitter (2001); Eisentraut (1956); Hill (1968); Hutterer and Joger (1982); Hutterer et al. (1992); Monadjem and Fahr (2007); Peterson and Smith (1973); Seemark and van Cakenberghe (2012); Thorn et al. (2007); Vielliard (1974); Hill (1983); Bergmans (1989); Bergmans (1990); Lunde et al. (2001); Schlitter et al. (1982); Dorst (1963); Aellen and Brosset (1968); Bergmans (1988); Brosset (1966a); Dowsett et al. (1991); Allen (1917); Ansell (1974); Bergmans (1997); Bergmans (1989); Fahr and Ebigo (2003); Fahr et al. (2002); Gallagher and Harrison (1977); Harrison (1975); Heller et al. (1994); Kearney and Seemark (2005); Kock and Howell (1988); Schouteden (1947); van Cakenberghe and de Vree (1985); Verschuren (1967); Jones (1971); Brosset (1966b); Csorba et al. (2003); and Malbrant and MaLatchy (1949). Records from West Africa: Robbins (1980); Green (1983); Koch-Weser (1984); Koopman et al. (1978); Kock (1969); Seemark and van Cakenberghe (2012); Decher and Fahr (2007); Grubb et al. (1998); Aellen (1956); Aellen (1963); Bergmans (1989); Eisentraut and Knorr (1957); Fahr and Ebigo (2003); Fahr et al. (2002); Fahr et al. (2006); Konstantinov et al. (2000); Kock (1969); Lopes and Crawford-Cabral (1992); Monard (1939); Bergmans et al. (1974); Cotterill (2001a); Cotterill (2001b); de Vree (1971); Fahr and Kalko (2011); Lim and Van Coeverden De Groot (1997); Coe (1975); Hill (1982); Koopman (1989); Koopman et al. (1995); Monadjem et al. (2013);

Monadjem et al. (In Press); Denys et al. (2013); Kuhn (1965); Monadjem and Fahr (2007); Monadjem (2011); Verschuren (1977); Wolton et al. (1982); Koopman (1975); Meining (2000); Sidiyene and Tranier (1990); Qumsiyeh and Schlitter (1981); Fairion (1980); Kock (1978); Poche (1975); Happold (1987); Hill et al. (1988); Adam and Hubert (1972); Adam and Hubert (1976); Adam et al. (1993); Bohme and Hutterer (1979); Poulet (1972); Decher et al. (2010); Weber and Fahr (2009); and Kock et al. (2002).

Species	Functional Group	Conservation Status	No. Records
<i>Asellia tridens</i>	Clutter	Least concern	26
<i>Cardioderma cor</i>	Clutter	Least concern	84
<i>Casinycteris argynnis</i>	Clutter	Least concern	11
<i>Chaerephon aloysiisabaudiae</i>	Open air	Least concern	9
<i>Chaerephon ansorgei</i>	Open air	Least concern	65
<i>Chaerephon bemmeleni</i>	Open air	Least concern	15
<i>Chaerephon bivittatus</i>	Open air	Least concern	51
<i>Chaerephon chapini</i>	Open air	Least concern	24
<i>Chaerephon major</i>	Open air	Least concern	46
<i>Chaerephon nigeriae</i>	Open air	Least concern	73
<i>Chaerephon pumilus</i>	Open air	Least concern	538
<i>Cistugo lesueuri</i>	Clutter edge	Least concern	19
<i>Cistugo seabrae</i>	Clutter edge	Least concern	14
<i>Cloeotis percivali</i>	Clutter	Least concern	36
<i>Coleura afra</i>	Open air	Least concern	66
<i>Eidolon helvum</i>	Clutter	Near threatened	333
<i>Epomophorus angolensis</i>	Clutter	Near threatened	21
<i>Epomophorus crypturus</i>	Clutter	Least concern	127
<i>Epomophorus gambianus</i>	Clutter	Least concern	176
<i>Epomophorus labiatus</i>	Clutter	Least concern	199
<i>Epomophorus minimus</i>	Clutter	Least concern	59
<i>Epomophorus wahlbergi</i>	Clutter	Least concern	249
<i>Epomops buettikoferi</i>	Clutter	Least concern	65
<i>Epomops dobsonii</i>	Clutter	Least concern	34
<i>Epomops franqueti</i>	Clutter	Least concern	174
<i>Eptesicus floweri</i>	Clutter edge	Least concern	8
<i>Eptesicus hottentotus</i>	Clutter edge	Least concern	51
<i>Glauconycteris alboguttata</i>	Clutter edge	Least concern	5
<i>Glauconycteris argentata</i>	Clutter edge	Least concern	38
<i>Glauconycteris beatrix</i>	Clutter edge	Least concern	16
<i>Glauconycteris humeralis</i>	Clutter edge	Data deficient	9
<i>Glauconycteris poensis</i>	Clutter edge	Least concern	34
<i>Glauconycteris variegata</i>	Clutter edge	Least concern	82

<i>Hipposideros abae</i>	Clutter	Least concern	42
<i>Hipposideros beatus</i>	Clutter	Least concern	53
<i>Hipposideros caffer</i>	Clutter	Least concern	540
<i>Hipposideros cyclops</i>	Clutter	Least concern	110
<i>Hipposideros fuliginosus</i>	Clutter	Least concern	30
<i>Hipposideros gigas</i>	Clutter	Least concern	74
<i>Hipposideros jonesi</i>	Clutter	Near threatened	25
<i>Hipposideros marisae</i>	Clutter	Vulnerable	8
<i>Hipposideros megalotis</i>	Clutter	Least concern	14
<i>Hipposideros ruber</i>	Clutter	Least concern	270
<i>Hipposideros vittatus</i>	Clutter	Near threatened	104
<i>Hypsignathus monstrosus</i>	Clutter	Least concern	146
<i>Hypsugo anchietae</i>	Clutter edge	Least concern	51
<i>Hypsugo bellieri</i>	Clutter edge	Least concern	7
<i>Hypsugo crassulus</i>	Clutter edge	Least concern	12
<i>Hypsugo eisentrauti</i>	Clutter edge	Data deficient	6
<i>Hypsugo musciculus</i>	Clutter edge	Data deficient	12
<i>Kerivoula argentata</i>	Clutter	Least concern	35
<i>Kerivoula cuprosa</i>	Clutter	Data deficient	5
<i>Kerivoula lanosa</i>	Clutter	Least concern	54
<i>Kerivoula phalaena</i>	Clutter	Least concern	13
<i>Kerivoula smithii</i>	Clutter	Least concern	15
<i>Laephotis angolensis</i>	Clutter edge	Data deficient	5
<i>Laephotis botswanae</i>	Clutter edge	Least concern	30
<i>Laephotis namibensis</i>	Clutter edge	Least concern	6
<i>Laephotis wintoni</i>	Clutter edge	Least concern	10
<i>Lavia frons</i>	Clutter	Least concern	254
<i>Lissonycteris angolensis</i>	Clutter	Least concern	174
<i>Megaloglossus woermannii</i>	Clutter	Least concern	105
<i>Micropteropus pusillus</i>	Clutter	Least concern	240
<i>Mimetillus moloneyi</i>	Clutter edge	Least concern	66
<i>Mimetillus thomasi</i>	Clutter edge	Not evaluated	11
<i>Miniopterus fraterculus</i>	Clutter edge	Least concern	26
<i>Miniopterus inflatus</i>	Clutter edge	Least concern	60
<i>Miniopterus minor</i>	Clutter edge	Data deficient	10
<i>Miniopterus natalensis</i>	Clutter edge	Least concern	313
<i>Miniopterus villiersi</i>	Clutter edge	Not evaluated	25
<i>Mops brachypterus</i>	Open air	Least concern	5
<i>Mops condylurus</i>	Open air	Least concern	279
<i>Mops demonstrator</i>	Open air	Not evaluated	25
<i>Mops leonis</i>	Open air	Least concern	29
<i>Mops midas</i>	Open air	Least concern	66
<i>Mops nanulus</i>	Open air	Least concern	32
<i>Mops niveiventer</i>	Open air	Least concern	32

<i>Mops spurrelli</i>	Open air	Least concern	24
<i>Mops thersites</i>	Open air	Data deficient	40
<i>Mops trevori</i>	Clutter edge	Vulnerable	10
<i>Myonycteris relictus</i>	Clutter edge	Least concern	5
<i>Myonycteris torquata</i>	Open air	Data deficient	78
<i>Myopterus daubentonii</i>	Open air	Least concern	7
<i>Myopterus whitelyi</i>	Clutter edge	Least concern	15
<i>Myotis bocagii</i>	Clutter edge	Vulnerable	127
<i>Myotis scotti</i>	Clutter edge	Least concern	9
<i>Myotis tricolor</i>	Clutter edge	Least concern	85
<i>Myotis welwitschii</i>	Clutter	Least concern	53
<i>Nanonycteris veldkampii</i>	Clutter edge	Near threatened	59
<i>Neoromicia brunnea</i>	Clutter edge	Least concern	15
<i>Neoromicia capensis</i>	Clutter edge	Data deficient	504
<i>Neoromicia cf. melckorum</i>	Clutter edge	Least concern	13
<i>Neoromicia guineensis</i>	Clutter edge	Data deficient	31
<i>Neoromicia helios</i>	Clutter edge	Least concern	15
<i>Neoromicia nana</i>	Clutter edge	Not evaluated	604
<i>Neoromicia pusillus</i>	Clutter edge	Least concern	5
<i>Neoromicia rendalli</i>	Clutter edge	Least concern	69
<i>Neoromicia somalica</i>	Clutter edge	Least concern	68
<i>Neoromicia tenuipinnis</i>	Clutter edge	Least concern	119
<i>Neoromicia airuluensis</i>	Clutter	Least concern	96
<i>Nycteris argo</i>	Clutter	Least concern	100
<i>Nycteris aurita</i>	Clutter	Least concern	21
<i>Nycteris gambiensis</i>	Clutter	Least concern	57
<i>Nycteris grandis</i>	Clutter	Least concern	88
<i>Nycteris hispida</i>	Clutter	Least concern	350
<i>Nycteris intermedia</i>	Clutter	Least concern	18
<i>Nycteris macrotis</i>	Clutter	Data deficient	234
<i>Nycteris major</i>	Clutter	Least concern	8
<i>Nycteris nana</i>	Clutter	Least concern	47
<i>Nycteris thebaica</i>	Clutter	Least concern	593
<i>Nycteris woodi</i>	Clutter edge	Least concern	27
<i>Nycticeinops schlieffeni</i>	Open air	Near threatened	243
<i>Otomops martiensseni</i>	Clutter edge	Data deficient	43
<i>Pipistrellus aero</i>	Clutter edge	Not evaluated	6
<i>Pipistrellus grandidieri</i>	Clutter edge	Not evaluated	26
<i>Pipistrellus hesperidus</i>	Clutter edge	Least concern	123
<i>Pipistrellus inexpectatus</i>	Clutter edge	Data deficient	7
<i>Pipistrellus nanulus</i>	Clutter edge	Least concern	41
<i>Pipistrellus rueppellii</i>	Clutter edge	Least concern	98
<i>Pipistrellus rusticus</i>	Clutter edge	Least concern	75
<i>Platymops setiger</i>	Open air	Least concern	19

<i>Plerotes anchietae</i>	Clutter	Data deficient	10
<i>Rhinolophus alcyone</i>	Clutter	Least concern	36
<i>Rhinolophus blasii</i>	Clutter	Least concern	61
<i>Rhinolophus capensis</i>	Clutter	Least concern	27
<i>Rhinolophus clivosus</i>	Clutter	Least concern	283
<i>Rhinolophus darlingi</i>	Clutter	Least concern	151
<i>Rhinolophus deckenii</i>	Clutter	Near threatened	12
<i>Rhinolophus denti</i>	Clutter	Least concern	31
<i>Rhinolophus eloquens</i>	Clutter	Least concern	48
<i>Rhinolophus fumigatus</i>	Clutter	Least concern	190
<i>Rhinolophus guineensis</i>	Clutter	Vulnerable	23
<i>Rhinolophus hildebrandtii</i>	Clutter	Least concern	214
<i>Rhinolophus hillorum</i>	Clutter	Near threatened	11
<i>Rhinolophus hipposideros</i>	Clutter	Least concern	5
<i>Rhinolophus landeri</i>	Clutter	Least concern	189
<i>Rhinolophus maclaudi</i>	Clutter	Endangered	8
<i>Rhinolophus ruwenzorii</i>	Clutter	Vulnerable	11
<i>Rhinolophus simulator</i>	Clutter	Least concern	120
<i>Rhinolophus swinnyi</i>	Clutter	Least concern	62
<i>Rhinopoma cystops</i>	Clutter	Least concern	20
<i>Rhinopoma macinnesi</i>	Open air	Data deficient	10
<i>Rhinopoma microphyllum</i>	Open air	Least concern	11
<i>Rousettus aegyptiacus</i>	Clutter	Least concern	245
<i>Rousettus lanosus</i>	Clutter	Least concern	45
<i>Saccopteryx peli</i>	Open air	Least concern	57
<i>Sauromys petrophilus</i>	Open air	Least concern	69
<i>Scotoecus albofuscus</i>	Clutter edge	Data deficient	17
<i>Scotoecus hindei</i>	Clutter edge	Least concern	54
<i>Scotoecus hirundo</i>	Clutter edge	Least concern	26
<i>Scotonycteris ophiodon</i>	Clutter	Near threatened	10
<i>Scotonycteris zenkeri</i>	Clutter	Least concern	44
<i>Scotophilus damarensis</i>	Clutter edge	Least concern	48
<i>Scotophilus dinganii</i>	Clutter edge	Least concern	464
<i>Scotophilus leucogaster</i>	Clutter edge	Least concern	96
<i>Scotophilus nigrita</i>	Clutter edge	Least concern	23
<i>Scotophilus nucella</i>	Clutter edge	Data deficient	5
<i>Scotophilus nux</i>	Clutter edge	Least concern	20
<i>Scotophilus viridis</i>	Clutter edge	Least concern	102
<i>Tadarida aegyptiaca</i>	Open air	Least concern	214
<i>Tadarida fulminans</i>	Open air	Least concern	33
<i>Tadarida lobata</i>	Open air	Least concern	9
<i>Tadarida ventralis</i>	Open air	Data deficient	20
<i>Taphozous hamiltoni</i>	Open air	Data deficient	11
<i>Taphozous hildegardeae</i>	Open air	Vulnerable	12
<i>Taphozous mauritianus</i>	Open air	Least concern	245

<i>Taphozous nudiventris</i>	Open air	Least concern	37
<i>Taphozous perforatus</i>	Open air	Least concern	79
<i>Triaenops afer</i>	Clutter edge	Least concern	50

Table A.2 AUC values indicate the discriminatory value of the models to predict the likelihood of a presence point being higher than the likelihood of a pseudo-absence point. The area under the curve (AUC) values for each climate change model used in the study. The climate change scenarios used were the current scenario, RCP2.6 (p26bi50), RCP4.5 (ip45bi50), RPC6.0 (ip60bi50), RPC8.5 (ip85bi50). Only models with values equal or higher than 0.7 were used in the study.

Species	AUC Current	AUC ip26bi50	AUC ip45bi50	AUC ip60bi50	AUC ip85bi50
<i>Asellia tridens</i>	0.86	0.86	0.86	0.86	0.86
<i>Cardioderma cor</i>	0.93	0.93	0.93	0.93	0.93
<i>Casinycteris argynnus</i>	0.96	0.96	0.96	0.96	0.96
<i>Chaerephon aloysiisabaudiae</i>	0.88	0.88	0.88	0.88	0.88
<i>Chaerephon ansorgei</i>	0.86	0.86	0.86	0.86	0.86
<i>Chaerephon bemmeleni</i>	0.89	0.89	0.89	0.89	0.89
<i>Chaerephon bivittatus</i>	0.88	0.88	0.88	0.88	0.88
<i>Chaerephon chapini</i>	0.82	0.82	0.82	0.82	0.82
<i>Chaerephon major</i>	0.87	0.87	0.87	0.87	0.87
<i>Chaerephon nigeriae</i>	0.85	0.85	0.85	0.85	0.85
<i>Chaerephon pumilus</i>	0.84	0.84	0.84	0.84	0.84
<i>Cistugo lesueuri</i>	0.99	0.99	0.99	0.99	0.99
<i>Cistugo seabrae</i>	0.95	0.95	0.95	0.95	0.95
<i>Cloeotis percivali</i>	0.93	0.93	0.93	0.93	0.93
<i>Coleura afra</i>	0.88	0.88	0.88	0.88	0.88
<i>Eidolon helvum</i>	0.80	0.80	0.80	0.80	0.80
<i>Epomophorus angolensis</i>	0.96	0.96	0.96	0.96	0.96
<i>Epomophorus crypturus</i>	0.94	0.94	0.94	0.94	0.94
<i>Epomophorus gambianus</i>	0.88	0.88	0.88	0.88	0.88
<i>Epomophorus labiatus</i>	0.92	0.92	0.92	0.92	0.92
<i>Epomophorus minimus</i>	0.92	0.92	0.92	0.92	0.92
<i>Epomophorus wahlbergi</i>	0.92	0.92	0.92	0.92	0.92
<i>Epomops buettikoferi</i>	0.98	0.98	0.98	0.98	0.98
<i>Epomops dobsonii</i>	0.97	0.97	0.97	0.97	0.97
<i>Epomops franqueti</i>	0.94	0.94	0.94	0.94	0.94
<i>Eptesicus floweri</i>	0.95	0.95	0.95	0.95	0.95
<i>Eptesicus hottentotus</i>	0.93	0.93	0.93	0.93	0.93

<i>Glauconycteris alboguttata</i>	0.92	0.92	0.92	0.92	0.92
<i>Glauconycteris argentata</i>	0.94	0.94	0.94	0.94	0.94
<i>Glauconycteris beatrix</i>	0.95	0.95	0.95	0.95	0.95
<i>Glauconycteris humeralis</i>	0.97	0.97	0.97	0.97	0.97
<i>Glauconycteris poensis</i>	0.95	0.95	0.95	0.95	0.95
<i>Glauconycteris variegata</i>	0.79	0.79	0.79	0.79	0.79
<i>Hipposideros abae</i>	0.95	0.95	0.95	0.95	0.95
<i>Hipposideros beatus</i>	0.95	0.95	0.95	0.95	0.95
<i>Hipposideros caffer</i>	0.83	0.83	0.83	0.83	0.83
<i>Hipposideros cyclops</i>	0.93	0.93	0.93	0.93	0.93
<i>Hipposideros fuliginosus</i>	0.94	0.94	0.94	0.94	0.94
<i>Hipposideros gigas</i>	0.90	0.90	0.90	0.90	0.90
<i>Hipposideros jonesi</i>	0.95	0.95	0.95	0.95	0.95
<i>Hipposideros megalotis</i>	0.97	0.97	0.97	0.97	0.97
<i>Hipposideros ruber</i>	0.87	0.87	0.87	0.87	0.87
<i>Hipposideros vittatus</i>	0.91	0.91	0.91	0.91	0.91
<i>Hypsipathus monstrosus</i>	0.93	0.93	0.93	0.93	0.93
<i>Hypsugo anchietae</i>	0.95	0.95	0.95	0.95	0.95
<i>Hypsugo bellieri</i>	0.95	0.95	0.95	0.95	0.95
<i>Hypsugo crassulus</i>	0.90	0.90	0.90	0.90	0.90
<i>Hypsugo eisentrauti</i>	0.91	0.91	0.91	0.91	0.91
<i>Hypsugo musciculus</i>	0.93	0.93	0.93	0.93	0.93
<i>Kerivoula argentata</i>	0.87	0.87	0.87	0.87	0.87
<i>Kerivoula cuprosa</i>	0.87	0.87	0.87	0.87	0.87
<i>Kerivoula lanosa</i>	0.88	0.88	0.88	0.88	0.88
<i>Kerivoula phalaena</i>	0.95	0.95	0.95	0.95	0.95
<i>Kerivoula smithii</i>	0.93	0.93	0.93	0.93	0.93
<i>Laephotis angolensis</i>	0.98	0.98	0.98	0.98	0.98
<i>Laephotis botswanae</i>	0.96	0.96	0.96	0.96	0.96
<i>Laephotis namibensis</i>	0.98	0.98	0.98	0.98	0.98
<i>Laephotis wintoni</i>	0.92	0.92	0.92	0.92	0.92
<i>Lavia frons</i>	0.89	0.89	0.89	0.89	0.89
<i>Lissonycteris angolensis</i>	0.92	0.92	0.92	0.92	0.92
<i>Megalochirus woermannii</i>	0.95	0.95	0.95	0.95	0.95
<i>Micropteropus pusillus</i>	0.90	0.90	0.90	0.90	0.90
<i>Mimetillus moloneyi</i>	0.93	0.93	0.93	0.93	0.93
<i>Mimetillus thomasi</i>	0.86	0.86	0.86	0.86	0.86
<i>Miniopterus fraterculus</i>	0.98	0.98	0.98	0.98	0.98
<i>Miniopterus inflatus</i>	0.90	0.90	0.90	0.90	0.90
<i>Miniopterus minor</i>	0.91	0.91	0.91	0.91	0.91
<i>Miniopterus natalensis</i>	0.91	0.91	0.91	0.91	0.91

<i>Miniopterus villiersi</i>	0.98	0.98	0.98	0.98	0.98
<i>Mops brachypterus</i>	0.87	0.87	0.87	0.87	0.87
<i>Mops condylurus</i>	0.97	0.97	0.97	0.97	0.97
<i>Mops demonstrator</i>	0.95	0.95	0.95	0.95	0.95
<i>Mops leonis</i>	0.84	0.84	0.84	0.84	0.84
<i>Mops midas</i>	0.91	0.91	0.91	0.91	0.91
<i>Mops nanulus</i>	0.94	0.94	0.94	0.94	0.94
<i>Mops niveiventer</i>	0.96	0.96	0.96	0.96	0.96
<i>Mops spurrelli</i>	0.94	0.94	0.94	0.94	0.94
<i>Mops thersites</i>	0.88	0.88	0.88	0.88	0.88
<i>Mops trevori</i>	0.70	0.70	0.70	0.70	0.70
<i>Myonycteris reducta</i>	0.95	0.95	0.95	0.95	0.95
<i>Myonycteris torquata</i>	0.82	0.82	0.82	0.82	0.82
<i>Myopterus daubentonii</i>	0.97	0.97	0.97	0.97	0.97
<i>Myopterus whitneyi</i>	0.85	0.85	0.85	0.85	0.85
<i>Myotis bocagii</i>	0.98	0.98	0.98	0.98	0.98
<i>Myotis scotti</i>	0.94	0.94	0.94	0.94	0.94
<i>Myotis tricolor</i>	0.93	0.93	0.93	0.93	0.93
<i>Myotis welwitschii</i>	0.96	0.96	0.96	0.96	0.96
<i>Nanonycteris veldkampii</i>	0.97	0.97	0.97	0.97	0.97
<i>Neoromicia airuluensis</i>	0.93	0.93	0.93	0.93	0.93
<i>Neoromicia brunnea</i>	0.89	0.89	0.89	0.89	0.89
<i>Neoromicia capensis</i>	0.95	0.94	0.95	0.95	0.95
<i>Neoromicia cf. melckorum</i>	0.85	0.85	0.85	0.85	0.85
<i>Neoromicia guineensis</i>	0.95	0.94	0.95	0.95	0.95
<i>Neoromicia helios</i>	0.84	0.84	0.84	0.84	0.84
<i>Neoromicia nana</i>	0.86	0.86	0.86	0.86	0.86
<i>Neoromicia pusillus</i>	0.79	0.78	0.79	0.79	0.79
<i>Neoromicia rendalli</i>	0.87	0.86	0.87	0.87	0.87
<i>Neoromicia somalica</i>	0.92	0.92	0.92	0.92	0.92
<i>Neoromicia tenuipinnis</i>	0.94	0.93	0.94	0.94	0.94
<i>Nycteris arge</i>	0.94	0.94	0.94	0.94	0.94
<i>Nycteris aurita</i>	0.92	0.92	0.92	0.92	0.92
<i>Nycteris gambiae</i>	0.92	0.92	0.92	0.92	0.92
<i>Nycteris grandis</i>	0.83	0.82	0.83	0.83	0.83
<i>Nycteris hispida</i>	0.89	0.90	0.89	0.89	0.89
<i>Nycteris intermedia</i>	0.81	0.81	0.81	0.81	0.81
<i>Nycteris macrotis</i>	0.89	0.89	0.89	0.89	0.89
<i>Nycteris major</i>	0.95	0.95	0.95	0.95	0.95
<i>Nycteris nana</i>	0.84	0.84	0.84	0.84	0.84
<i>Nycteris thebaica</i>	0.96	0.96	0.96	0.96	0.96

<i>Nycterus woodi</i>	0.87	0.87	0.87	0.87	0.87
<i>Nycticeinops schlieffeni</i>	0.88	0.88	0.88	0.88	0.88
<i>Otomops martiensseni</i>	0.94	0.94	0.94	0.94	0.94
<i>Pipistrellus aero</i>	0.97	0.97	0.97	0.97	0.97
<i>Pipistrellus grandidieri</i>	0.74	0.71	0.74	0.74	0.74
<i>Pipistrellus hesperidus</i>	0.94	0.94	0.94	0.94	0.94
<i>Pipistrellus inexpectatus</i>	0.90	0.90	0.90	0.90	0.90
<i>Pipistrellus nanulus</i>	0.90	0.89	0.90	0.90	0.90
<i>Pipistrellus rueppellii</i>	0.74	0.76	0.74	0.74	0.74
<i>Pipistrellus rusticus</i>	0.87	0.87	0.87	0.87	0.87
<i>Platymops setiger</i>	0.95	0.95	0.95	0.95	0.95
<i>Plerotes anchietae</i>	0.98	0.98	0.98	0.98	0.98
<i>Rhinolophus alcyone</i>	0.94	0.94	0.94	0.94	0.94
<i>Rhinolophus blasii</i>	0.94	0.94	0.94	0.94	0.94
<i>Rhinolophus capensis</i>	0.99	0.99	0.99	0.99	0.99
<i>Rhinolophus clivosus</i>	0.93	0.93	0.93	0.93	0.93
<i>Rhinolophus darlingi</i>	0.94	0.94	0.94	0.94	0.94
<i>Rhinolophus deckenii</i>	0.91	0.91	0.91	0.91	0.91
<i>Rhinolophus denti</i>	0.94	0.93	0.94	0.94	0.94
<i>Rhinolophus eloquens</i>	0.96	0.96	0.96	0.96	0.96
<i>Rhinolophus fumigatus</i>	0.86	0.85	0.86	0.86	0.86
<i>Rhinolophus guineensis</i>	0.97	0.97	0.97	0.97	0.97
<i>Rhinolophus hildebrandtii</i>	0.92	0.92	0.92	0.92	0.92
<i>Rhinolophus hillorum</i>	0.99	0.99	0.99	0.99	0.99
<i>Rhinolophus hipposideros</i>	0.70	0.70	0.70	0.70	0.70
<i>Rhinolophus landeri</i>	0.79	0.79	0.79	0.79	0.79
<i>Rhinolophus maclaudi</i>	0.96	0.96	0.96	0.96	0.96
<i>Rhinolophus ruwenzorii</i>	0.99	0.99	0.99	0.99	0.99
<i>Rhinolophus simulator</i>	0.94	0.94	0.94	0.94	0.94
<i>Rhinolophus swinnyi</i>	0.95	0.95	0.95	0.95	0.95
<i>Rhinopoma cystops</i>	0.90	0.91	0.90	0.90	0.90
<i>Rhinopoma macinnesi</i>	0.90	0.90	0.90	0.90	0.90
<i>Rhinopoma microphyllum</i>	0.83	0.83	0.83	0.83	0.83
<i>Rousettus aegyptiacus</i>	0.89	0.89	0.89	0.89	0.89
<i>Rousettus lanosus</i>	0.98	0.98	0.98	0.98	0.98
<i>Saccopteryx peli</i>	0.95	0.95	0.95	0.95	0.95
<i>Sauromys petrophilus</i>	0.94	0.95	0.94	0.94	0.94
<i>Scotoecus albofuscus</i>	0.84	0.80	0.84	0.84	0.84
<i>Scotoecus hindei</i>	0.83	0.84	0.83	0.83	0.83
<i>Scotoecus hirundo</i>	0.87	0.88	0.87	0.87	0.87
<i>Scotonycteris ophiodon</i>	0.93	0.93	0.93	0.93	0.93

<i>Scotonycteris zenkeri</i>	0.96	0.96	0.96	0.96	0.96
<i>Scotophilus damarensis</i>	0.97	0.97	0.97	0.97	0.97
<i>Scotophilus dinganii</i>	0.86	0.86	0.86	0.86	0.86
<i>Scotophilus leucogaster</i>	0.88	0.88	0.88	0.88	0.88
<i>Scotophilus nigrita</i>	0.86	0.84	0.86	0.86	0.86
<i>Scotophilus nucella</i>	0.97	0.97	0.97	0.97	0.97
<i>Scotophilus nux</i>	0.97	0.97	0.97	0.97	0.97
<i>Scotophilus viridis</i>	0.90	0.90	0.90	0.90	0.90
<i>Tadarida aegyptiaca</i>	0.91	0.90	0.91	0.91	0.91
<i>Tadarida fulminans</i>	0.89	0.91	0.89	0.89	0.89
<i>Tadarida lobata</i>	0.73	0.73	0.73	0.73	0.73
<i>Tadarida ventralis</i>	0.81	0.81	0.81	0.81	0.81
<i>Taphozous hamiltoni</i>	0.99	0.99	0.99	0.99	0.99
<i>Taphozous hildegardeae</i>	0.93	0.93	0.93	0.93	0.93
<i>Taphozous mauritianus</i>	0.82	0.83	0.82	0.82	0.82
<i>Taphozous nudiventris</i>	0.85	0.86	0.85	0.85	0.85
<i>Taphozous perforatus</i>	0.79	0.80	0.79	0.79	0.79
<i>Triaenops afer</i>	0.84	0.84	0.84	0.84	0.84

Table A.3 The original land system category from van Asselen & Verburg, 2013) and the land use categories used in the analyses to simplify the data

Land System	Land Use Category
Cropland extensive, few livestock	Intensive and extensive agriculture
Cropland extensive, bovines, goats and sheep	Intensive and extensive agriculture
Cropland extensive, pigs and poultry	Intensive and extensive agriculture
Cropland medium intensive, few livestock	Intensive and extensive agriculture
Cropland medium intensive, bovines, goats and sheep	Intensive and extensive agriculture
Cropland medium intensive, pigs and poultry	Intensive and extensive agriculture
Cropland intensive, few livestock	Intensive and extensive agriculture
Cropland intensive, bovines, goats and sheep	Intensive and extensive agriculture
Cropland intensive, pigs and poultry	Intensive and extensive agriculture
Mosaic cropland and grassland, bovines, goats and sheep	Agricultural and natural mosaic
Mosaic cropland and grassland, pigs and poultry	Agricultural and natural mosaic
Mosaic cropland (ext.) and grassland, few livestock	Agricultural and natural mosaic
Mosaic cropland (med. Int.) and grassland, few livestock	Agricultural and natural mosaic
Mosaic cropland (int.) and grassland, few livestock	Agricultural and natural mosaic
Mosaic cropland and forest	Agricultural and natural mosaic
Mosaic cropland (ext.) and forest, few livestock	Agricultural and natural mosaic
Mosaic cropland (med. Int.) and forest, few livestock	Agricultural and natural mosaic
Mosaic cropland (int.) and forest, few livestock	Agricultural and natural mosaic
Dense forest	Forest and low intensity agriculture
Forest, few livestock	Forest and low intensity agriculture
Forest, pigs and poultry	Forest and low intensity agriculture
Mosaic grassland and forest	Grassland and low intensity agriculture
Mosaic grassland and bare	Grassland and low intensity agriculture
Grassland, natural	Grassland and low intensity agriculture
Grassland, few livestock	Grassland and low intensity agriculture
Grassland, bovines, goats and sheep	Grassland and low intensity agriculture
Bare	Bare ground
Bare, goats and sheep	Bare ground
Peri-urban and villages	Urban
Urban	Urban

Table A.4 Condition values for each land use used in the Zonation analysis under and optimistic and pessimistic scenarios

		Bat Functional Group Open Air	Bat Functional Group Clutter Edge	Bat Functional Group Clutter
Optimistic	Agricultural and natural mosaic	0.80	0.75	0.70
	Bare ground	0.19	0.04	0.13
	Forest and low intensity agriculture	1.00	1.00	1.00
	Grassland and low intensity agriculture	0.70	0.50	0.40
	Intensive and extensive agriculture	0.65	0.60	0.50
	Urban	0.60	0.55	0.50

		Bat Functional Group Open Air	Bat Functional Group Clutter Edge	Bat Functional Group Clutter
Pessimistic	Agricultural and natural mosaic	0.50	0.45	0.40
	Bare ground	0.19	0.04	0.13
	Forest and low intensity agriculture	1.00	1.00	1.00
	Grassland and low intensity agriculture	0.50	0.30	0.20
	Intensive and extensive agriculture	0.35	0.30	0.20
	Urban	0.40	0.35	0.30

Figure A.1 Points indicating the locations of bat records used in the analyses

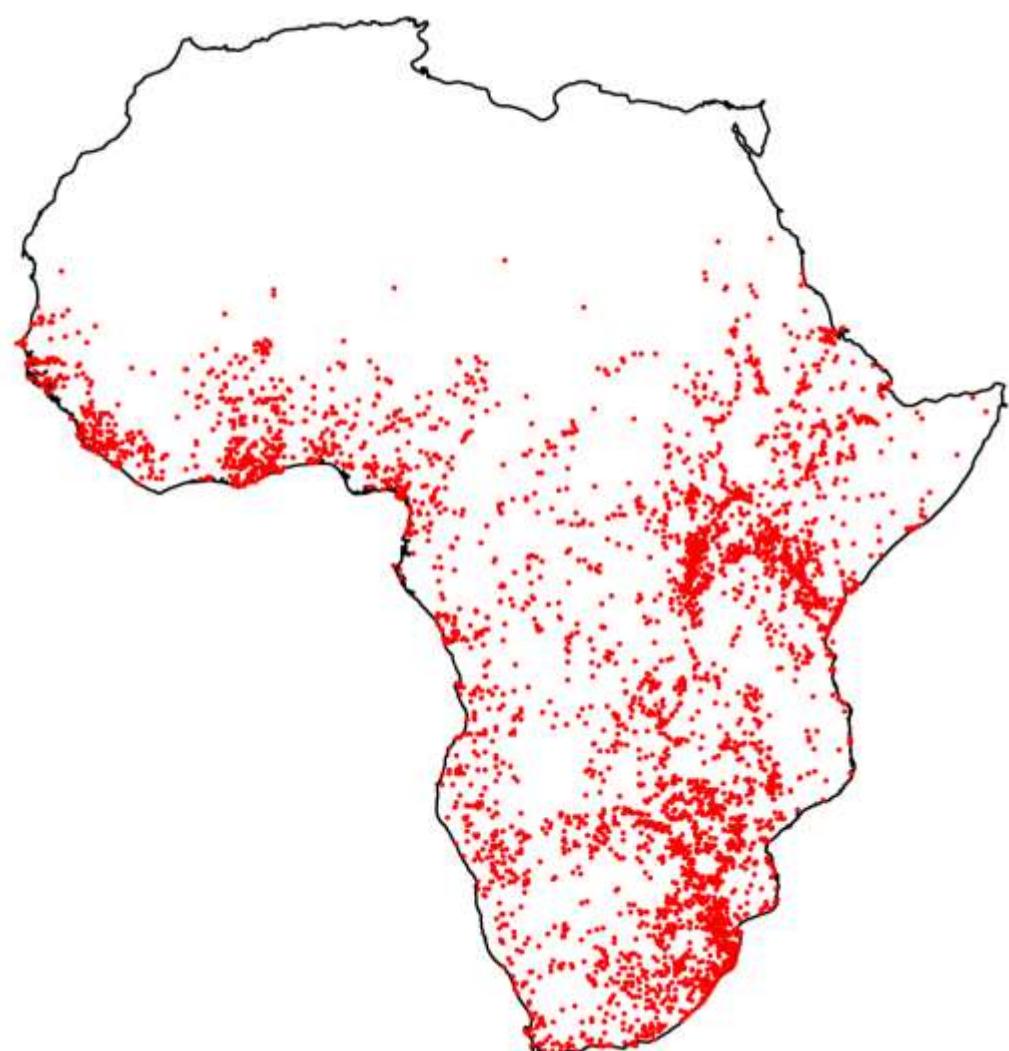


Figure A.2 Flow chart showing the order of analyses and inputs used in the Zonation prioritization analysis.

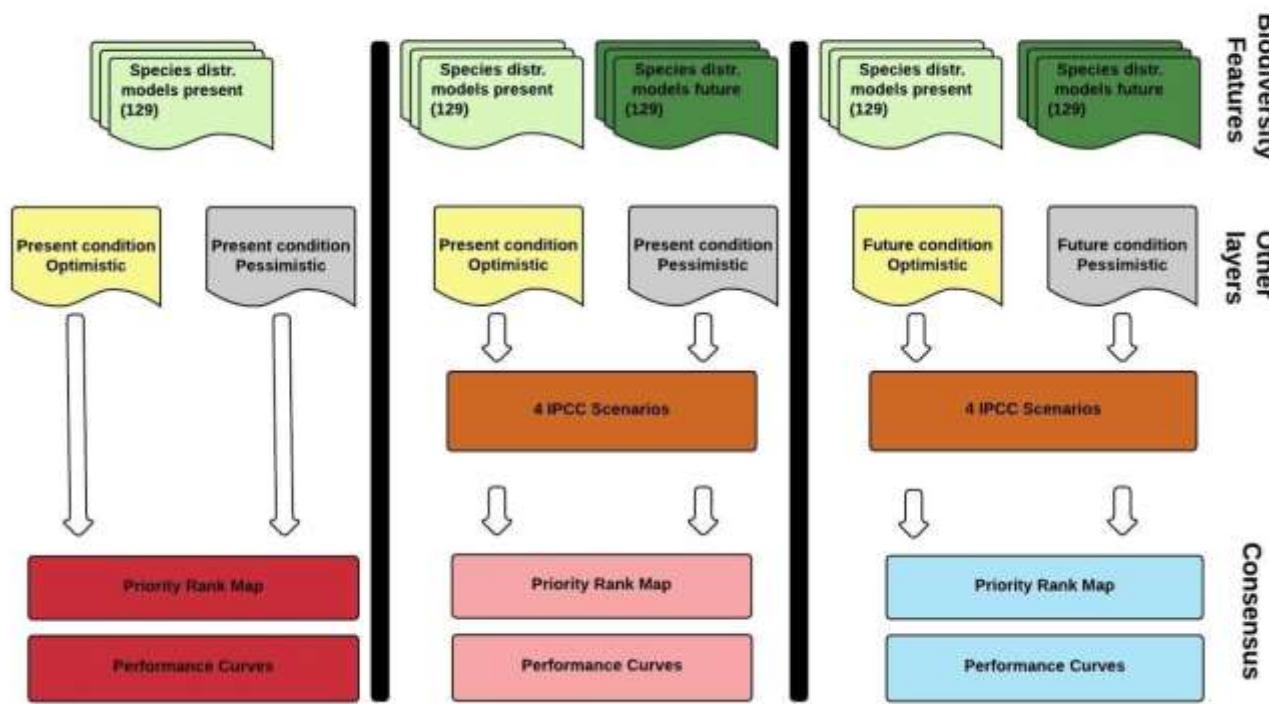


Figure A.3 The present and future land use coverages used in the analyses.

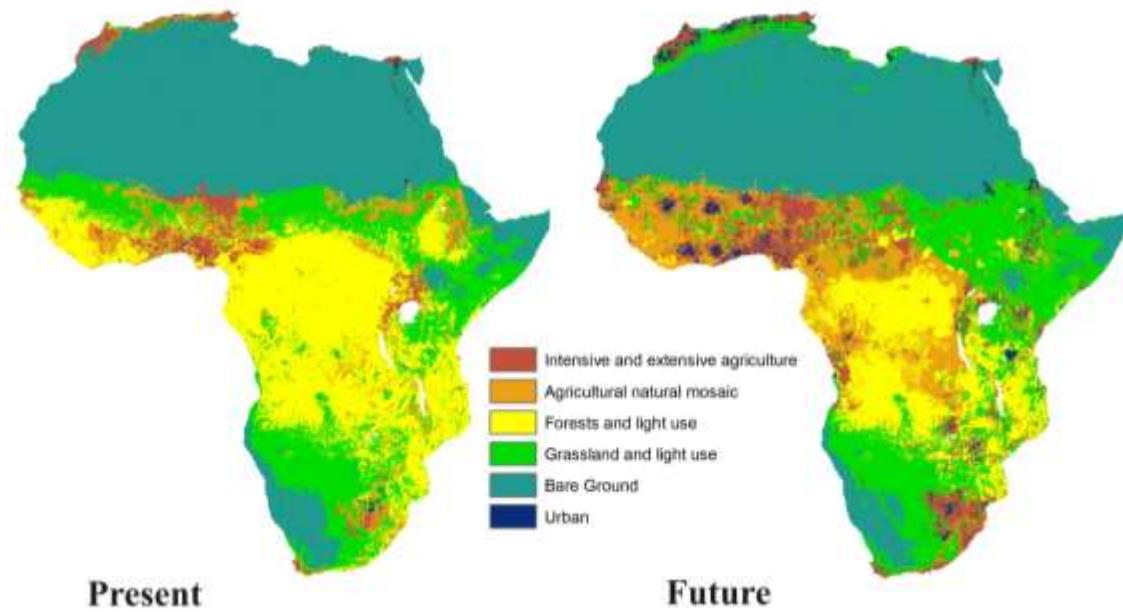
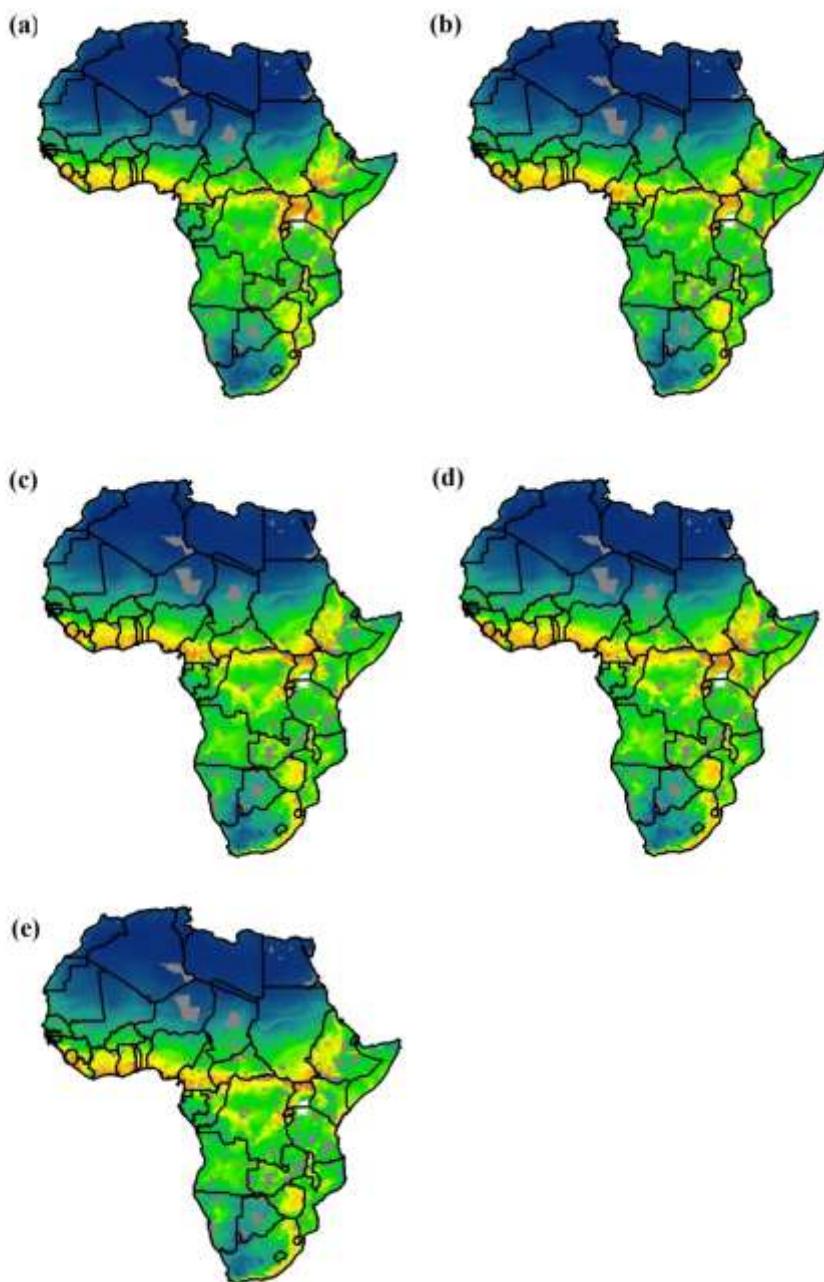


Figure A.4 The summed probabilities of MaxEnt predicted occurrence for species used in the analyses.

Scenarios for current distribution: a. current prediction using current climatic conditions to predict distribution
future distribution: b. predicted future distribution using RCP2.6, c. RCP4.5 d. RPC6.0 and e. RPC8.5
scenarios.



Appendix B

Supplementary Results

Table B.1 The percentage overlap between the 17 percent conservation core areas of each scenario used in the study. CCCO (Current climate and current optimistic), CCCP (Current climate and current pessimistic), FCCO (Future climate change current pessimistic land use), FCFO (Future climate and future optimistic land use), FCFP (Future climate and future pessimistic land use).

	CCCO	CCCP	FCCO	FCCP	FCFO	FCFP
CCCO	100	85	82	81	67	58
CCCP	85	100	80	82	67	59
FCCO	82	80	100	84	67	58
FCCP	81	82	84	100	68	59
FCFO	67	67	67	68	100	75
FCFP	58	59	58	59	75	100

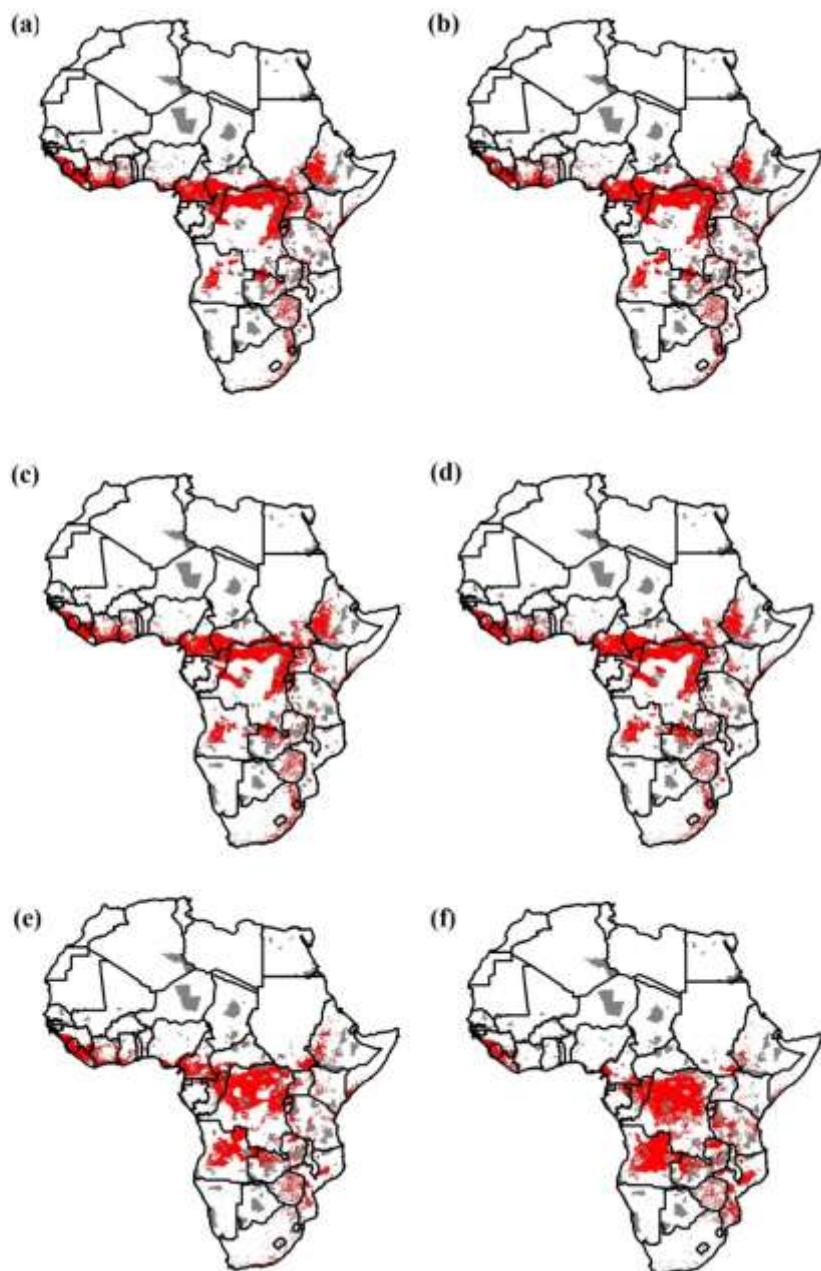
Table B.2 The average, minimum and maximum proportion of species ranges falling within current protected areas.

	Current	Future Climate Change Only	Future Climate Change and Land Use Change
Average Representation	0.069	0.068	0.065
Minimum Representation	0.02	0.015	0.019
Maximum Representation	0.12	0.13	0.12

Table B.3 Median representation for different IUCN categories of bats at the 17% priority areas. CCCO (current climate and current optimistic land use), CCCP (current climate and current pessimistic land use), FCCO (future climate and current optimistic land use), FCPO (future climate and current pessimistic land use), FCFO (future climate and future optimistic land use), FCFP (future climate future and pessimistic land use). None of the species included in the analysis is critically endangered.

	LC	TH	VU/DD
CCCO	0.283	0.306	0.299
CCCP	0.269	0.291	0.273
FCCO	0.328	0.352	0.349
FCCO	0.348	0.398	0.362
FCFO	0.295	0.329	0.333
FCFP	0.286	0.305	0.330

Figure B.1 Conservation priority areas within the 17 percent conservation target (red areas) and current conserved areas (grey); (a) current climate and current optimistic land use (b) current climate and current pessimistic land use (c) future climate and current optimistic land use (d) future climate and current pessimistic land use (e) future climate and future optimistic land use (f) future climate future and future pessimistic land use.



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